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# Densities and Habitat of American Brook Lamprey (Lampetra appendix) Larvae in Minnesota

NEAL D. MUNDAHL,  $^1$  GOLAM SAYEED, STEPHEN TAUBEL, CONRAD ERICKSON, ANDREW ZALATEL AND JULIE COUSINS

Department of Biology and Large River Studies Center, Winona State University, P.O. Box 5838, Winona, Minnesota 55987

ABSTRACT.—At least one formerly large population of American brook lamprey (Lampetra appendix) has been extirpated in Minnesota, but little is known about the species in the state. We examined densities and age structures of American brook lamprey larvae in several streams in southeastern Minnesota. Mean densities of lamprey larvae in the best habitats available in 13 streams varied from 0.33–5.78 larvae/m². Seven of nine streams examined had missing age classes of larvae, with five streams missing two or more classes. Habitat use and length—weight relationships of larvae were similar to those reported previously for this and other species of lamprey. Many L. appendix populations in Minnesota are healthy, but others are susceptible to extirpation because of low densities and multiple missing age classes. The current status of many of these populations, and their susceptibility to further disturbance, need to be more thoroughly examined to better assess the need for legal protection for the species in Minnesota.

#### INTRODUCTION

The American brook lamprey Lampetra appendix (DeKay) has the widest geographic distribution of any nonparasitic lamprey in North America (Page and Burr, 1991), but populations have declined or been eliminated from many streams because of human activities that impact the lamprey and its stream habitats (Renaud, 1997; Mundahl, 1998). Residential development within drainage basins (Eddy and Underhill, 1974), dam construction and flood-control projects (Trautman, 1981), pollution from agricultural activities and logging (Eddy and Underhill, 1974; Trautman, 1981), bait collecting (Vladykov, 1949, 1973) and sea lamprey Petromyzon marinus control projects (Manion and Purvis, 1971; Cochran et al., 1993) all have been detrimental to populations of L. appendix. Populations on the periphery of the species' range are most threatened (NatureServe, 2003), and several states in these regions (e.g., Connecticut, Massachusetts, Vermont, Iowa) have heeded pleas (Vladykov, 1973; Renaud, 1997) to protect North American nonparasitic lampreys in affording L. appendix protected status (Johnson, 1987; Hoff, 1988; Schmidt, 1995).

In Minnesota, the American brook lamprey is restricted primarily to the southeastern portion of the state (Phillips et al., 1982), where streams have been impacted by human activities for over a century (Waters, 1977; Thorn et al., 1997) and where several populations of Lampetra appendix have been extirpated (Eddy and Underhill, 1974; Phillips et al., 1982). Minnesota lies on the northwestern periphery of the range of L. appendix. Although American brook lamprey was listed as a Special Concern species in Minnesota prior to 1996, it was removed from the list after the species was collected in a variety of streams in the southeastern part of the state during the early- and mid-1990s. However, no comprehensive

<sup>&</sup>lt;sup>1</sup> Corresponding author: Telephone: (507)457-5695; e-mail: nmundahl@winona.edu

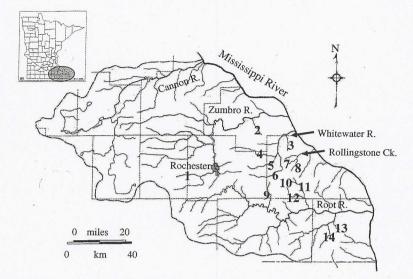


Fig. 1.—Map of streams in southeastern Minnesota, with major drainages and streams where American brook lamprey larvae were collected. Inset highlights location of study area in Minnesota. Study streams were: 1 – Salem Creek, 2 – West Indian Creek, 3 – Trout Valley Creek, 4 – North Branch Whitewater River, 5 – Middle Branch Whitewater River, 6 – South Branch Whitewater River, 7 – Bear Creek, 8 – Rupprecht Creek, 9 – Trout Run Creek, 10 – Ferguson Creek, 11 – Rush Creek, 12 – Pine Creek, 13 – Badger Creek, 14 – Beaver Creek

surveys, or even general population assessments, of L. appendix have ever been conducted within this region.

During a series of stream surveys, watershed assessments and student research projects conducted over a period of 5 y, we collected a variety of data on American brook lamprey larvae and their habitats in many streams throughout southeastern Minnesota. Here we summarize these data in an attempt to determine the present status of several populations of American brook lamprey within four drainages in southeastern Minnesota and to evaluate the decision to not provide state-wide protection for the species. We examined densities and age structures of larvae to assess population status and also evaluated their length-weight patterns and habitat use for comparison to previous work on the species in other portions of its range.

## STUDY AREA

Streams in southeastern Minnesota extend across two ecoregions (Driftless Area, Western Corn Belt Plains; Omernik and Gallant, 1988) and include a mix of coldwater, coolwater and warmwater systems. Four major drainages (Cannon River – 3785 km², Zumbro River – 3656 km², Root River – 4275 km², Whitewater River – 758 km²) and many other smaller systems all are tributary to the Mississippi River (Fig. 1). Land use is dominated by agriculture (mixed row crop, hay and pasture), although several areas support managed forests or moderate-sized (30,000–75,000 people) urban areas. This region of rolling uplands and steeply sloped valleys was subjected to intensive agricultural development and logging in the late-1800s, resulting in increased erosion, flooding and stream sedimentation and decreased discharges of coldwater springs (Surber, 1924; Waters, 1977; Trimble and Lund,

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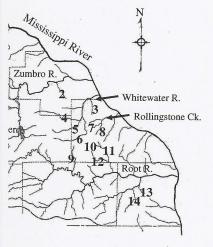
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1982; Thorn *et al.*, 1997). Deterioration of stream habitat quality and overexploitation led to extirpation of some of the native fish fauna, most notably brook trout *Salvelinus fontinalis*, in many streams by 1900 (Thorn *et al.*, 1997).

Since 1930, stream habitat degradation in southeastern Minnesota has diminished and some natural stream rehabilitation has occurred as land conservation practices have been implemented (Thorn *et al.*, 1997). Much instream habitat today remains degraded, but extensive sections of stream habitat have been improved for trout (mostly introduced brown trout *Salmo trutta*) over the past several decades (Thorn *et al.*, 1997), with approximately 5 km of new habitat improvements added each year [Minnesota Department of Natural Resources (MDNR), 1998]. Trout management activities in this region today involve 1145 km of 139 streams (Thorn *et al.*, 1997).

This study focused on a five-county area where *Lampetra appendix* was most common in historical collections (MDNR, 1995). Other lampreys (northern brook lamprey *Ichthyomyzon fossor* and silver lamprey *I. unicuspis*) are extremely rare within this region (Phillips *et al.*, 1982).

#### MATERIALS AND METHODS

During various research and assessment projects conducted in southeastern Minnesota between April and September, 1995–1999, American brook lamprey larvae were collected from their burrows in fine sediments with the aid of backpack electrofishing units (Smith Root Type VII and 12-B POW). Based on prior collecting experience, habitats judged most suitable and likely to contain lamprey larvae were electrofished intensively and all emerging larvae were captured until no additional specimens emerged from the sediments in three successive passes. The total area sampled was measured to determine larval density (Beamish and Lowartz, 1996). This approach was used to determine densities of larvae in the best habitat available, rather than to assess an overall density for an entire stream segment. No attempt was made to determine the total area of habitat suitable for lamprey larvae in each stream.

Larvae collected from several streams were anesthetized (MS-222), weighed (nearest 0.1 g wet weight), measured [mm total length (TL)] and returned to their habitat after recovery. Size measurements were used to assess length-weight patterns of larvae and population age structure (based on smoothed length-frequency distributions; Hardisty and Potter, 1971; Beamish and Medland, 1988) at each of the stream sites. Populations examined during summer were expected to contain a maximum of five size/age groups of larvae, whereas those examined during autumn through spring should contain four size/age groups of larvae plus metamorphosed, prespawning adults (Seagle and Nagel, 1982). Although statoliths can be used to determine American brook lamprey age and verify age classes (e.g., Beamish and Medland, 1988), they were not used because of a no-kill policy in effect during the study period.

During some of the lamprey surveys conducted during late-summer and autumn (baseflow stream conditions present), habitat assessments were made at the point of collection of individual larvae. Water depth and average water column current velocity [cm/s at 0.6-depth (60% of distance from stream surface to bottom); Marsh-McBirney Flowmate current meter] were measured and a sample (>100 ml) of the substrate (top 5 cm) from which the larvae emerged was collected and returned to the laboratory for particle size (serial sieving of dried sediments through 0.2-, 0.4- and 0.8-mm mesh sieves) and volatile organic matter (weight loss on ignition of dried sediments at 550 C for 3 h) analyses.

TABLE 1.—Data collected on American brook lamprey larvae and their habitats in 14 streams within four drainages of the upper Mississippi River in southeastern Minnesota. X indicates data present

Drainage/stream	Density	Age classes	Length-weight	Habitat
Whitewater River drainage				
N. Br. Whitewater	X	X	X	X
M. Br. Whitewater	X	X	X	
S. Br. Whitewater	X	X	X	
Trout Valley Creek	X	X	X	X
Root River drainage				
Rush Creek	X			
Beaver Creek	X	X	X	X
Badger Creek	X	X	X	
Trout Run Creek	X			
Pine Creek	X	X	X	X
Ferguson Creek				X
Rollingstone Creek drainage				
Bear Creek	'X		X	X
Rupprecht Creek	X		X	
Zumbro River drainage				
Salem Creek	X	X	X	X
W. Indian Creek	X	X	X	X

Because of the varying goals of the different projects for which data on lamprey larvae were gathered, comparable data were not available for all of the 14 streams examined in this study. Density estimates were conducted in 13 streams, age structures were examined in nine streams, length-weight relationships were assessed in 11 streams and habitat data were gathered from eight streams (Table 1).

## RESULTS

## DENSITY ESTIMATES, AGE STRUCTURES AND LENGTH-WEIGHT PATTERNS OF LAMPREY LARVAE

Densities of American brook lamprey larvae were determined for small areas (n = 85, mean =  $4.22~\text{m}^2$ , range = 0.12– $19.60~\text{m}^2$ ) of habitat in 13 streams within four drainage basins (Fig. 1). Mean densities ranged from 0.33 larvae/m² in Trout Run Creek to 5.78 larvae/m² in Pine Creek (Table 2), and differed significantly (Kruskal-Wallis test: H= 30.89, P < 0.001) among streams. However, since densities were quite variable among streams within the same drainage (Table 2), there was no significant (Kruskal-Wallis test: H= 5.34, P = 0.15) difference in densities among the four drainages examined. Mean density in a stream was not correlated with either the length range of larvae collected (simple linear regression: F = 1.10, P = 0.32, r² = 0.08) or the number of age groups of larvae present (simple linear regression: F = 0.30, P = 0.60, r² = 0.04). Overall mean density for all streams was 3.57 larvae/m². Maximum densities found in any single location within each stream ranged from 0.33 to 16.67 larvae/m², with maximum densities averaging 5.4 larvae/m² across all streams.

Lamprey population age structures (*i.e.*, number and size of age classes present) based on length-frequencies varied widely among streams for both April and August samples (Table 2). Small, age-1 larvae comprised the most abundant age class in most (eight of 11) samples during both periods. In April 1997 and 1998, when four age classes of larvae were expected,

Table 2.—Mean ( $\pm sd$ ) densities (number/n in 13 streams within four drainages of the upp

Drainage/stream	Number
Whitewater River drainage	
N. Br. Whitewater	
M. Br. Whitewater	
S. Br. Whitewater	
Trout Valley Creek	2
Root River drainage	
Rush Creek	
Beaver Creek	
Badger Creek	
Trout Run Creek	
Pine Creek	1
Rollingstone Creek drainage	
Bear Creek	
Rupprecht Creek	
Zumbro River drainage	
Salem Creek	
W. Indian Creek	

only two of five streams contained all fou missing intermediate-size (ages class 2 or 4 larvae. Dominant year classes had been s streams). In August 1999, when five age

Table 3.—Year classes of American brook la during April 1997–1998 and August 1999. Va Year class 0 larvae were not available for captu sample size

Stream	N	0
April 1997		
N. Br. Whitewater	56	
Pine Creek	48	
April 1998		
Salem Creek	77	
Trout Valley Creek	109	
West Indian Creek	64	
August 1999		
N. Br. Whitewater	124	
M. Br. Whitewater	78	
S. Br. Whitewater	37	
Trout Valley Creek	62	3
Beaver Creek	20	
Badger Creek	22	

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rey larvae and their habitats in 14 streams within heastern Minnesota. X indicates data present

e classes	Length-weight	Habitat
X X	X	X
X	X	
X X	X	
X	X	X
X	X	X
X X	X	
X	X	X
		X
	X	X
	X	21
X	X	X
X	X Translation	
	Λ	X

projects for which data on lamprey larvae e for all of the 14 streams examined in this streams, age structures were examined in assessed in 11 streams and habitat data

## H-WEIGHT PATTERNS OF LAMPREY LARVAE

vere determined for small areas (n = 85, t in 13 streams within four drainage basins m² in Trout Run Creek to 5.78 larvae/m² (Kruskal-Wallis test: H = 30.89, P < 0.001) te variable among streams within the same Kruskal-Wallis test: H = 5.34, P = 0.15) examined. Mean density in a stream was ae collected (simple linear regression: F = groups of larvae present (simple linear ill mean density for all streams was 3.57 : location within each stream ranged from weraging 5.4 larvae/m<sup>2</sup> across all streams. r and size of age classes present) based on or both April and August samples (Table nt age class in most (eight of 11) samples 1 four age classes of larvae were expected,

Table 2.—Mean ( $\pm sd$ ) densities (number/m²) and density ranges of American brook lamprey larvae in 13 streams within four drainages of the upper Mississippi River in southeastern Minnesota

Drainage/stream	Number of sites	Density	Range	
Whitewater River drainage				
N. Br. Whitewater	9	3.63 (1.35)	1.70-5.13	
M. Br. Whitewater	2	4.70 (2.40)	3.00-6.40	
S. Br. Whitewater	4	2.52 (1.00)	1.71-3.96	
Trout Valley Creek	23	5.74 (3.83)	0.75 - 16.00	
Root River drainage				
Rush Creek	5	2.10 (0.95)	0.57-3.00	
Beaver Creek	4	1.17 (0.75)	0.44 - 2.00	
Badger Creek	4	3.69 (2.88)	2.00-8.00	
Trout Run Creek	2	0.33 (0.00)	0.33-0.33	
Pine Creek	11	5.78 (4.59)	1.16-16.67	
Rollingstone Creek drainage				
Bear Creek	4	1.04 (0.45)	0.56-1.50	
Rupprecht Creek	3	0.55 (0.39)	0.33 - 1.00	
Zumbro River drainage				
Salem Creek	6	1.93 (1.11)	0.51 - 3.03	
W. Indian Creek	. 8	1.79 (1.18)	0.34-3.81	

only two of five streams contained all four age groups of larvae (Table 3). Two streams were missing intermediate-size (ages class 2 or 3) larvae and one stream contained no large, age-4 larvae. Dominant year classes had been spawned in either 1996 (three streams) or 1997 (two streams). In August 1999, when five age classes of larvae may have been present (although

TABLE 3.—Year classes of American brook lamprey larvae present in southeastern Minnesota streams during April 1997–1998 and August 1999. Values are percentages of all individuals captured at a site. Year class 0 larvae were not available for capture in April, but should have been available in August. N is sample size

					Year class	e i - 127	7.74.6
Stream	N		0	1	2	3	4
April 1997							
N. Br. Whitewater	56			80	13		7
Pine Creek	48			46	29	21	4
April 1998							
Salem Creek	77		-	21	52	22	5
Trout Valley Creek	109			61	36	3	
West Indian Creek	64			88		2	10
August 1999		1					
N. Br. Whitewater	124			83	14		3
M. Br. Whitewater	78			92	8		
S. Br. Whitewater	37			5	3	14	78
Trout Valley Creek	62		3	58	3		
Beaver Creek	20						100
Badger Creek	22			41	59		

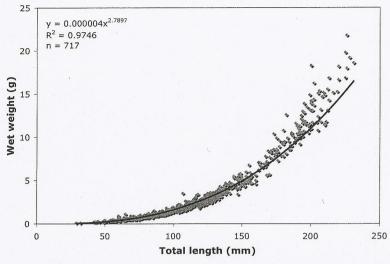


Fig. 2.—Total length-wet weight relationship of American brook lamprey larvae collected from 11 streams in southeastern Minnesota, 1997–1999. See Table 1 for the list of streams included

small, age-O larvae are difficult to collect), only one of six streams contained 4 age classes (Table 3). Three streams were missing both ages 3 and 4 larvae, whereas one stream contained only age 4 larvae. Dominant age classes had been spawned in 1995 (two streams), 1997 (one stream) or 1998 (three streams).

Length-weight relationships of lamprey larvae were determined for specimens from 11 different streams (Table 1). Larvae ranged from 29–231 mm TL and 0.1–21.8 g wet weight. Larvae from most streams exhibited similar length-weight patterns, but significant differences in these relationships (analysis of covariance: F = 23.79, P < 0.01) were detected among streams. However, these differences were attributed mostly to two streams (Bear and Beaver creeks), where only small numbers of very large larvae were collected. Since these differences were judged to be minor, measurements of larvae from all 11 streams were combined to produce a single length-weight relationship (Fig. 2).

## ASSESSMENTS OF LARVAL LAMPREY HABITATS

American brook lamprey larvae typically were collected from soft sediments in shallow ( $<50\,\mathrm{cm}$ ), slow-moving (5 cm/s) water (Table 4). However, larvae in some streams occupied habitats different from those in other streams. For example, mean water depths at larvae collection sites differed significantly (Kruskal-Wallis test: H=36.76, P <0.001) among streams, with Trout Valley Creek and North Branch Whitewater River collections generally coming from shallower sites (mean depths  $<30\,\mathrm{cm}$ ). Average water column current velocities also differed significantly (Kruskal-Wallis test: H=34.01, P <0.001) among streams, with Trout Valley Creek collection site velocities averaging 3–4× greater than those found at other streams. Depth and velocity at collection sites were not correlated (r=-0.13). Organic contents of surface sediments where larvae were collected averaged <10% (Table 4), but differed significantly (Kruskal-Wallis: H=40.82, P <0.001) among streams. Trout Valley Creek and West Indian Creek sediments had significantly lower organic contents (means <5%) than did sediments in other streams. However, there was no significant

TABLE 4.—Mean water depth, water colum surface sediments at locations where Americas southeastern Minnesota streams. Ranges are s

		Water d	
Stream	n	(cm	
Pine Creek	41	38 (15-	
N. Br. Whitewater R.	27	28 (11-	
W. Indian Creek	15	46 (20-	
Trout Valley Creek	21	21 (3-5	
Salem Creek	15	35 (17-	
Beaver Creek	4	49 (29-	
Ferguson Creek	3	50 (34-	
Bear Creek	1	20	

relationship (simple linear regression: al and either water depth, current veloci examined.

Sediments where larvae were collected by medium and fine sand and silt (particle gravel (>0.8 mm) were present in sampl abundance in most samples. Particle size-(contingency table:  $X^2 = 1727.1$ , P << 0 Trout Valley and West Indian creeks do

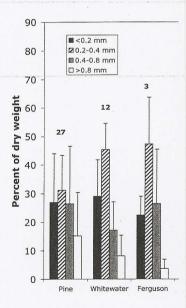
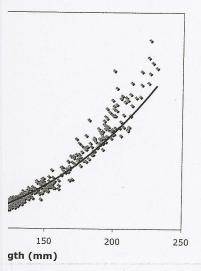


Fig. 3.—Particle size distributions of sedimer eight southeastern Minnesota streams. Bars a and numbers above bars are sample sizes

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American brook lamprey larvae collected from e Table 1 for the list of streams included

y one of six streams contained 4 age classes ages 3 and 4 larvae, whereas one stream es had been spawned in 1995 (two streams),

te were determined for specimens from 11 29–231 mm TL and 0.1–21.8 g wet weight. length-weight patterns, but significant covariance:  $F=23.79,\ P<0.01)$  were deces were attributed mostly to two streams imbers of very large larvae were collected. measurements of larvae from all 11 streams the relationship (Fig. 2).

## LAMPREY HABITATS

e collected from soft sediments in shallow . However, larvae in some streams occupied For example, mean water depths at larvae Vallis test: H=36.76, P<0.001) among ach Whitewater River collections generally 30 cm). Average water column current allis test: H=34.01, P<0.001) among elocities averaging 3–4× greater than those ection sites were not correlated (r=-0.13). vae were collected averaged <10% (Table = 40.82, P<0.001) among streams. Trout had significantly lower organic contents reams. However, there was no significant

TABLE 4.—Mean water depth, water column current velocity (0.6 depth) and organic content of surface sediments at locations where American brook lamprey larvae emerged from sediments in eight southeastern Minnesota streams. Ranges are shown in parentheses

Stream	n	Water depth (cm)	Current velocity (cm/s)	Sediment organic content (%)
Pine Creek	41	38 (15–100)	5 (0-20)	9.88 (1.73–18.54)
N. Br. Whitewater R.	27	28 (11-45)	5 (0-20)	10.97 (4.80-21.00)
W. Indian Creek	15	46 (20-75)	6 (2–11)	4.22 (2.19-6.60)
Trout Valley Creek	21	21 (3-52)	18 (0-42)	2.82 (0.51-4.80)
Salem Creek	15	35 (17-52)	4 (1–11)	6.50 (3.94–11.20)
Beaver Creek	4	49 (29-67)	4 (4-4)	8.20 (4.18–11.72)
Ferguson Creek	3	50 (34-62)	9 (5–14)	6.84 (4.20-8.52)
Bear Creek	1	20	0	6.01

relationship (simple linear regression: all P > 0.10) between the density of lamprey larvae and either water depth, current velocity or sediment organic content in the streams examined.

Sediments where larvae were collected most often were dominated (on a dry weight basis) by medium and fine sand and silt (particles <0.4 mm diameter; Fig. 3). Very coarse sand and gravel (>0.8 mm) were present in samples collected from all streams, but were very low in abundance in most samples. Particle size-class composition of sediments varied significantly (contingency table:  $X^2 = 1727.1$ , P << 0.001) among the streams, with sediments in Bear, Trout Valley and West Indian creeks dominated by much finer particles than were the

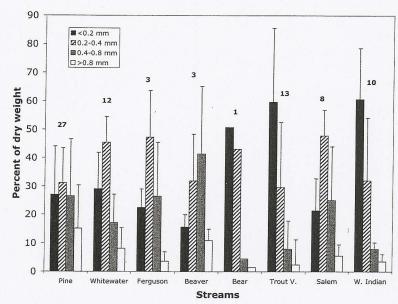


Fig. 3.—Particle size distributions of sediments from American brook lamprey larva collection sites in eight southeastern Minnesota streams. Bars are means, thin lines represent one standard deviation, and numbers above bars are sample sizes

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sediments collected from other streams. However, densities of lamprey larvae were no different (t-test: t=0.16, P=0.88) in these streams than in those with coarser sediments.

#### DISCUSSION

Densities of American brook lamprey larvae varied dramatically within and among streams sampled, despite selecting sites with habitats judged most suitable for larvae. Varying densities in southeastern Minnesota streams are not unusual for this (Beamish and Lowartz, 1996) or other lamprey species (Manion and McLain, 1971; Potter *et al.*, 1986; Beamish and Jebbink, 1994; Pajos and Weise, 1994). Much of the variation in densities can be attributed to slight variations in habitat, especially with respect to sediment particle size composition (*see* review by Beamish and Lowartz, 1996). The abundance of American brook lamprey larvae increases exponentially with increases in medium-fine sand (0.15–0.25 mm diam; Beamish and Lowartz, 1996), with larvae avoiding finer and coarser sediments (Anderson and White, 1988; Lee, 1989). However, in the present study, densities of larvae displayed no pattern related to sediment size composition. Densities of some species of lamprey larvae also may be affected by sediment organic content, current velocity and water depth, but none of these have been found to have a significant effect on densities of *Lampetra appendix* larvae (Anderson and White, 1988; Beamish and Lowartz, 1996; present study).

Very few studies have assessed population sizes or densities of American brook lamprey larvae (Smith and McLain, 1962; Waters et al., 1993; Simonson and Lyons, 1995; Beamish and Lowartz, 1996). Population estimates of Lampetra appendix larvae that included all stream habitats (Smith and McLain, 1962; Waters et al., 1993; Simonson and Lyons, 1995) are not comparable to densities measured only in suitable habitats as in the present study. Similarly, mean lamprey densities from the present study are not comparable to those reported by Beamish and Lowartz (1996), since the latter study included habitats where larvae were least likely to exist. However, maximum densities from the 13 streams in the present study (median =  $3.81 \text{ larvae/m}^2$ , mean =  $5.40 \text{ larvae/m}^2$ , range =  $0.33-16.67 \text{ larvae/m}^2$ m<sup>2</sup>) are comparable, but considerably lower, than maximum densities (median = 11.8 larvae/ $m^2$ , mean = 14.7 larvae/ $m^2$ , range = 3.7–26.7 larvae/ $m^2$ ) observed in 10 streams by Beamish and Lowartz (1996). In fact, nearly half (46%; 6 of 13 streams) of the Minnesota streams had maximum densities lower than the lowest stream maximum of Beamish and Lowartz (1996). Lower maximum densities can be indicative of reduced population sizes, poorer quality habitat or both. However, without information on the total amount of habitat available for lamprey larvae in each stream, it is not possible to estimate the impact of these lower maximum densities on the overall populations of L. appendix within these streams. Lower densities of larvae scattered throughout a more extensive habitat can compensate for the lack of a few, isolated groups of larvae at high densities.

Despite good densities of American brook lamprey larvae in many southeastern Minnesota streams, and the lack of a relationship between larval density and the number of age classes present in a stream, the reduced age structures observed at six of the nine stream sites examined is of concern. Missing age classes have been observed previously in populations of Lampetra appendix and other species of lamprey (Schaffner, 1902; Okkelberg, 1922; Hubbs, 1925; Thomas, 1962; Morman, 1979; Trautman, 1981; Seagle and Nagel, 1982). There is no evidence that American brook lamprey larvae segregate by size among different sites (Beamish and Lowartz, 1996), so missing age classes are not an artifact of sampling location. Missing or weak year classes in fish populations are common and can result from a wide array of environmental and biological causes (Scalet et al., 1996; Ross, 1997). It is possible that missing year classes are characteristic of, and possibly normal for, populations of American brook lamprey on the periphery of their geographic range.

The reduction or loss of a single age should not be a major problem and, if Lampetra appendix is semelparous with individuals requiring >5 y to reach mat Seagle and Nagel, 1982), and these may older year classes of lamprey larvae may metamorphosis (Morkert et al., 1998). He are present (e.g., Beaver Creek and Mi lamprey populations likely are at risk o flooding, interfere with spring spawning larvae. The flood-prone nature of streams brief (<1 wk) spawning period of Amescenario. For example, lamprey larvae we 1999, but following several years with sevin 2004 despite intensive sampling at sev

Degraded habitat also can be a cause of brook lamprey. Adult American brook lat spawning (Cochran et al., 1993). Lack of stexample, spawning substrates in Ruppred Only during years when sufficient numb displace these sediments will spawning blamprey spawners and successful spawning only once in 5 y, with generally only two observed in most years.

American brook lamprey larvae in Minidentical to those reported for two other Let al., 1976) and Ontario (Lanteigne et al. those exhibited by larvae of other lampraepyptera; Manion and McLain, 1971; Rollifestyle of these organisms limits the degrand populations.

The status of the American brook lamp though declines and losses have occurred center of its range, it is the most abunda Morman, 1979). Recent surveys in Wisconsi widespread and abundant than previously populations avoiding detection until recent Cochran, pers. comm.). Records indicate t nine drainages in Minnesota (MDNR, 199 American brook lamprey populations are present lack of public awareness (Mundahl a wide geographic area tend to support th from protected status in Minnesota in 1996. of L. appendix has disappeared from a Minne potentially at risk because of low densities habitat threats and spawning and/or early d problems, and the species' limited capacity of reduced or extirpated populations. Ba available (e.g., Beamish and Lowartz, 1996

Iowever, densities of lamprey larvae were no streams than in those with coarser sediments.

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arvae varied dramatically within and among nabitats judged most suitable for larvae. Varying are not unusual for this (Beamish and Lowartz, McLain, 1971; Potter et al., 1986; Beamish and of the variation in densities can be attributed respect to sediment particle size composition. The abundance of American brook lampreys in medium-fine sand (0.15–0.25 mm diam; oiding finer and coarser sediments (Anderson present study, densities of larvae displayed no 1. Densities of some species of lamprey larvae ontent, current velocity and water depth, but ificant effect on densities of Lampetra appendix and Lowartz, 1996; present study).

sizes or densities of American brook lamprey l., 1993; Simonson and Lyons, 1995; Beamish of Lampetra appendix larvae that included all aters et al., 1993; Simonson and Lyons, 1995) ly in suitable habitats as in the present study. present study are not comparable to those nce the latter study included habitats where ximum densities from the 13 streams in the  $n = 5.40 \text{ larvae/m}^2$ , range = 0.33-16.67 larvae/ r, than maximum densities (median = 11.8 .7-26.7 larvae/m<sup>2</sup>) observed in 10 streams by half (46%; 6 of 13 streams) of the Minnesota the lowest stream maximum of Beamish and an be indicative of reduced population sizes, ut information on the total amount of habitat is not possible to estimate the impact of these ulations of L. appendix within these streams. a more extensive habitat can compensate for high densities.

ook lamprey larvae in many southeastern ship between larval density and the number ed age structures observed at six of the nine age classes have been observed previously in ecies of lamprey (Schaffner, 1902; Okkelberg, , 1979; Trautman, 1981; Seagle and Nagel, ook lamprey larvae segregate by size among so missing age classes are not an artifact of es in fish populations are common and can d biological causes (Scalet *et al.*, 1996; Ross, re characteristic of, and possibly normal for, e periphery of their geographic range.

The reduction or loss of a single age class in a population of American brook lamprey should not be a major problem and, in most cases, should be only temporary. Although Lampetra appendix is semelparous with a 5-y life cycle (Okkelberg, 1922; Hubbs, 1925), individuals requiring >5 y to reach maturity are present in many systems (Thomas, 1962; Seagle and Nagel, 1982), and these may spawn in place of a missing year class. In addition, older year classes of lamprey larvae may be missing in streams with rapid growth and early metamorphosis (Morkert et al., 1998). However, in streams where only one or two age classes are present (e.g., Beaver Creek and Middle Branch Whitewater River), American brook lamprey populations likely are at risk of being lost if environmental conditions, such as flooding, interfere with spring spawning and/or development of eggs and newly hatched larvae. The flood-prone nature of streams in southeastern Minnesota (Waters, 1977) and the brief (<1 wk) spawning period of American brook lamprey makes this a very possible scenario. For example, lamprey larvae were common in South Branch Whitewater River in 1999, but following several years with severe prolonged floods, no lamprey were collected in 2004 despite intensive sampling at several sites.

Degraded habitat also can be a cause of missing age classes in populations of American brook lamprey. Adult American brook lamprey require riffle areas with clean substrate for spawning (Cochran et al., 1993). Lack of such areas will severely impair spawning efforts. For example, spawning substrates in Rupprecht Creek are buried beneath >5 cm of fine silt. Only during years when sufficient numbers of spawners are present to dig through and displace these sediments will spawning be successful. Large numbers of American brook lamprey spawners and successful spawning have been observed in this stream by the authors only once in 5 y, with generally only two or three adults and no active spawning nests observed in most years.

American brook lamprey larvae in Minnesota exhibited length-weight patterns virtually identical to those reported for two other *Lampetra appendix* populations in Delaware (Rohde et al., 1976) and Ontario (Lanteigne et al., 1981). These patterns also were very similar to those exhibited by larvae of other lamprey species (sea lamprey, least brook lamprey *L. aepyptera*; Manion and McLain, 1971; Rohde et al., 1976), indicating that the burrowing lifestyle of these organisms limits the degree of allometric variation possible among species and populations.

The status of the American brook lamprey in North America is apparently secure, even though declines and losses have occurred in some regions (NatureServe, 2003). In the center of its range, it is the most abundant and widely distributed lamprey species (e.g., Morman, 1979). Recent surveys in Wisconsin suggest that Lampetra appendix may be far more widespread and abundant than previously believed (Lyons et al., 2000), with many small populations avoiding detection until recent advances in stream electrofishing equipment (P. Cochran, pers. comm.). Records indicate that the species has been found in 46 streams in nine drainages in Minnesota (MDNR, 1995; Hatch et al., 2003), and many of the state's American brook lamprey populations are healthy, despite past habitat threats and the present lack of public awareness (Mundahl, 1998). Good populations in many streams over a wide geographic area tend to support the decision to remove American brook lamprey from protected status in Minnesota in 1996. However, at least one formerly large population of L. appendix has disappeared from a Minnesota stream (Phillips et al., 1982), and others are potentially at risk because of low densities of larvae and missing age classes. Continuing habitat threats and spawning and/or early development failures are the likely causes of these problems, and the species' limited capacity for dispersal (Morman, 1979) inhibits recovery of reduced or extirpated populations. Based on what little comparative information is available (e.g., Beamish and Lowartz, 1996), more information is needed on L. appendix,

especially on the periphery of its range in Minnesota, to adequately document current populations and assess their susceptibility to human activities. Additionally, with our knowledge of the habitat needs of American brook lamprey larvae, the amount of suitable habitat in area streams should be quantified to determine their ability to support viable populations of lamprey. Only after such information is in hand can it be decided whether or not the American brook lamprey should receive protected status in Minnesota.

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## Spatial and Tempo Darter Abundance North-ce

PETER C. SMILEY, JR., <sup>1</sup> ERIC D. I Forest and Wildlife Research Center, Mississippi

ABSTRACT.—Little is known about the habitat use of the goldstripe darter (Ethe species associates and habitat use of this pine plantations in north-central Missis among sampling periods differed among of goldstripe darters exhibited significan while the remaining 10 streams did r Goldstripe darters exhibited the great culatus), brown madtom (Noturus phae Goldstripe darter abundance was positiv and sand. Our results suggest that man forested riparian zones adjacent to first goldstripe darters.

The goldstripe darter (Etheostoma parasubgenus Fuscatelum (Bart and Taylor, 19 by the lack of bright breeding colors, the line (Bart and Taylor, 1999). Goldstripe from eastern Texas to central Georgia extent in southeastern Missouri and southeastern fuscated within its distribution across the southeast this species is considered rare or endanged Arkansas, Missouri) and the eastern-most Pigg, 1999; Arkansas Natural Heritage Cheritage Program, 2003; Herbert and Georgia and Taylor (1998).

Understanding the ecology and habit developing recovery plans designed to Unfortunately, information on the ecologistism of its range concerning the habitat use of the goldst contained in regional ichthyological surcaptured from small springs and spring

<sup>&</sup>lt;sup>1</sup> Corresponding author present address: Hayes Drive, Columbus, Ohio 43210. Teleph 50@osu.edu

<sup>&</sup>lt;sup>2</sup> Present address: Department of Forest Corvallis Oregon 97331